


RESEARCH

Open Access



The relevance of using in situ carbon and nitrogen data and satellite images to assess aboveground carbon and nitrogen stocks for supporting national REDD + programmes in Africa

Adéyèmi Chabi^{1,2*} , Sven Lautenbach^{3,4}, Jérôme Ebagnerin Tondoh², Vincent Oladokoun Agnila Orekan⁵, Stephen Adu-Bredu⁶, Nicholas Kyei-Baffour⁷, Vincent Joseph Mama⁸ and John Fonweban⁹

Abstract

Background: To reduce the uncertainty in estimates of carbon emissions resulting from deforestation and forest degradation, better information on the carbon density per land use/land cover (LULC) class and in situ carbon and nitrogen data is needed. This allows a better representation of the spatial distribution of carbon and nitrogen stocks across LULC. The aim of this study was to emphasize the relevance of using in situ carbon and nitrogen content of the main tree species of the site when quantifying the aboveground carbon and nitrogen stocks in the context of carbon accounting. This paper contributes to that, by combining satellite images with in situ carbon and nitrogen content in dry matter of stem woods together with locally derived and published allometric models to estimate aboveground carbon and nitrogen stocks at the Dassari Basin in the Sudan Savannah zone in the Republic of Benin.

Results: The estimated mean carbon content per tree species varied from $44.28 \pm 0.21\%$ to $49.43 \pm 0.27\%$. The overall mean carbon content in dry matter for the 277 wood samples of the 18 main tree species of the region was $47.01 \pm 0.28\%$ —which is close to the Tier 1 coefficient of 47% default value suggested by the Intergovernmental Panel on Climate Change (IPCC). The overall mean fraction of nitrogen in dry matter was estimated as $0.229 \pm 0.016\%$. The estimated mean carbon density varied from $1.52 \pm 0.14 \text{ Mg C ha}^{-1}$ (for Cropland and Fallow) to $97.83 \pm 27.55 \text{ Mg C ha}^{-1}$ (for *Eucalyptus grandis* Plantation). In the same order the estimated mean nitrogen density varied from $0.008 \pm 0.007 \text{ Mg ha}^{-1}$ of N (for Cropland and Fallow) to $0.321 \pm 0.088 \text{ Mg ha}^{-1}$ of N (for *Eucalyptus grandis* Plantation).

Conclusion: The results show the relevance of using the in situ carbon and nitrogen content of the main tree species for estimating aboveground carbon and nitrogen stocks in the Sudan Savannah environment. The results provide crucial information for carbon accounting programmes related to the implementation of the REDD+ initiatives in developing countries.

Keywords: Relevance, In situ, Carbon, Nitrogen, Assess, Aboveground, REDD + programmes, Africa, Sudan Savannah

*Correspondence: princechabi@gmail.com; chabi.a@wascal.org

² West African Science Service Centre On Climate Change and Adapted Land Use (WASCAL), Competence Centre Ouagadougou, 06 BP 9507 Ouaga 06 Ouagadougou, Burkina Faso

Full list of author information is available at the end of the article



Background

In the context of climate change issues, emissions from deforestation and forest degradation in developing countries constitute some 20 percent of the total global emission of greenhouse gases annually [1]. Thus, reducing emissions from deforestation and degradation, biodiversity conservation, sustainable forest management and enhancement of forest carbon stocks (REDD+) in developing countries has become important frameworks to mitigate climate change and limit the rise in global temperature to no more than 2 °C [1–3]. Current challenges for the management of forests and other land use classes are the development of verifiable, reliable, accurate and cost-effective methods to adequately document forest resources dynamics [2]. The estimation of aboveground carbon stocks and the related uncertainties arise from inadequate data [3, 4]. These uncertainties in turn compromise the estimation of terrestrial carbon emissions as well as the knowledge of in situ data [3, 5–7]. Better assessments of aboveground nitrogen stocks could also be of interest since they provide necessary information for predicting nitrous oxide emission from damaged or burned trees. The accuracy of the estimation of mean carbon and nitrogen density for each land use/land cover class depends thereby on reliable carbon and nitrogen content estimates per main tree species, species frequency estimates per land use/land cover class and the availability of reliable allometric models to infer oven-dry aboveground biomass of trees from tree census data [8].

Allometric equations have been used by many authors all over the world [8–19, 58, 59] for estimating biomass stocks of ecosystems. The estimation of carbon stocks in Sub-Saharan Africa is based on allometric models and forest inventory data [8, 20–31]. Many studies so far focused on the estimation of aboveground biomass of forest ecosystems, specific tree species or plantations [8, 20, 22, 23, 25, 27, 32–40, 60]. The study from Kuya [29] was few of them which focused on the estimation of aboveground biomass in agricultural landscapes. However, woody vegetation in agricultural landscapes represents a significant carbon pool. In sub-Saharan Africa, the majority (87%) of agriculturally dominated landscapes has a tree cover of more than 10% [41].

To reduce the uncertainty in estimates of carbon dioxide and nitrous oxide emissions from deforestation and forest degradation, more complete and higher quality information-based satellite images and in situ data is needed. The estimation of the total carbon and nitrogen stocks at the landscape level is complex since the vegetation pattern changes from one land use/land cover class to another and the tree species distribution varies gradually by size and species. Additionally, there is a need for reliable methods that are applicable to target species in

the region of interest [41]. With increasing data requirements and analytical complexity from Tier 1 to Tier 3, the accuracy and precision of the carbon estimate also increases [42]. An accurate estimation of aboveground carbon and nitrogen stocks is recommended by the IPCC [42] to considerably reduce the uncertainty in the Tier 3 approach. The Tier 1 approach [42] suggested a coefficient of 0.47 to convert mean biomass density to the mean carbon density for a defined ecosystem or land use/land cover class. This default value is applied in many cases at the national level by many developing countries in the absence of information on carbon content of the main tree species of the region. In some cases a coefficient of 0.5 has been applied [4, 43]. Both default values may underestimate or overestimate the carbon stock, leading to a substantial level of uncertainty. In addition to information on regional land use, specific conversion factors and allometric models are needed that allow a biomass estimation at the landscape scale based on properties that are easy and reliable to measure under field conditions. Conversion factors and allometric models can then be used together with remote sensing based land use/land cover information to estimate the current carbon and nitrogen stocks or to quantify the changes in these stocks.

The aim of this study was to quantify the aboveground carbon and nitrogen stocks at the landscape level for the current (2013–2014) land use/land cover at the scale of a watershed in the West African Sudan Savannah using in situ carbon and nitrogen content of the main tree species of the site.

Results and discussion

Carbon and nitrogen content of dry matter of the main tree species

The fraction of carbon and nitrogen in the dry matter of the wood samples of the main tree species of the Dassari watershed in this Sudan Savannah environment differed clearly between the different tree species (Table 1, Fig. 1). The tree species with a high mean carbon fraction were *Terminalia macroptera* ($49.43 \pm 0.24\%$), *Pterocarpus erinaceus* ($49.43 \pm 0.27\%$) and *Crotopteryx febrifuga* ($49.17 \pm 0.21\%$). The lowest carbon content of dry matter was obtained for *Combretum glutinosum* (min 41.73%) with the mean of the species of $44.72 \pm 0.44\%$ and the highest for *Acacia seyal* (max 53.07%) with the mean of the species of $46.50 \pm 0.68\%$. The estimated mean per tree species varied from $44.28 \pm 0.21\%$ to $49.43 \pm 0.27\%$. The overall mean of the 277 stem wood samples for all species was $47.01 \pm 0.28\%$ which is in line with default value of 47% in the IPCC [42] Tier 1 approach. The substantial variation of carbon content across tree species confirmed

Table 1 Carbon (C) and nitrogen (N) contents of stem wood of the main tree species of the watershed

Trees species	n	Carbon (C) contents (% dm)			Nitrogen (N) contents (% dm)			DBH (cm)		C/N ratio		
		Min	Max	Mean (SE)	Min	Max	Mean (SE)	Min	Max	Min	Max	Mean (SE)
<i>Terminalia macroptera</i>	19	46.267	51.241	49.474 (0.266)	0.108	0.303	0.192 (0.013)	9.3	40.7	160.50	428.39	281.81 (18.33)
<i>Terminalia avicennioides</i>	03	47.971	49.759	48.70 (0.53)	0.155	0.181	0.168 (0.007)	16.6	24	265.03	312.16	289.96 (13.67)
<i>Acacia seyal</i>	14	43.928	53.071	46.50 (0.684)	0.13	0.583	0.290 (0.037)	7.6	34.4	80.71	357.6	194.24 (21.72)
<i>Acacia gourmaensis</i>	02	47.55	48.09	47.824 (0.269)	0.297	0.349	0.323 (0.025)	13.4	19	160.11	137.80	148.95 (11.15)
<i>Combretum glutinosum</i>	11	41.737	45.959	44.72 (0.438)	0.14	0.358	0.241 (0.020)	8	32	125.94	320.95	201.36 (19.15)
<i>Pterocarpus erinaceus</i>	21	46.779	51.645	49.438 (0.278)	0.164	0.427	0.242 (0.014)	6.9	44.7	110.09	295.09	216.28 (10.63)
<i>Anogeisus leiocarpus</i>	16	44.037	46.003	44.917 (0.167)	0.08	0.273	0.128 (0.012)	6.9	32.4	161.30	570.05	386.52 (28.28)
<i>Mitragyna inermis</i>	18	44.978	47.74	46.724 (0.174)	0.177	0.354	0.243 (0.011)	7	34.5	129.46	262.19	199.40 (9.23)
<i>Lannea microcarpa</i>	20	42.091	45.938	44.282 (0.209)	0.148	0.405	0.273 (0.015)	7	50.3	110.95	306.08	173.47 (11.14)
<i>Lannea acida</i>	6	43.408	45.164	44.526 (0.248)	0.14	0.386	0.265 (0.035)	10.8	36	115.60	320.80	186.92 (30.61)
<i>Ficus sp</i>	21	43.931	46.38	45.153 (0.139)	0.16	0.427	0.294 (0.015)	8.6	52.7	105.3	286.90	163.14 (9.83)
<i>Crotopteryx febrifuga</i>	18	47.662	52.229	49.172 (0.217)	0.118	0.306	0.182 (0.014)	5.6	30.6	161.14	417.54	295.68 (20.50)
<i>Entada Africana</i>	15	45.852	48.377	47.098 (0.191)	0.242	0.475	0.357 (0.016)	8.4	27.6	100.09	196.18	135.97 (6.75)
<i>Parkia biglobosa</i>	23	44.02	47.636	46.516 (0.214)	0.127	0.396	0.201 (0.013)	8.6	62.4	119.40	358.43	247.85 (12.35)
<i>Vitellaria paradoxa</i>	22	45.972	50.032	47.942 (0.228)	0.13	0.337	0.228 (0.010)	8	60	136.41	367.23	220.11 (11.37)
<i>Azadirachta indica</i>	16	47.253	52.999	49.005 (0.413)	0.104	0.302	0.177 (0.014)	8.8	50.5	162.43	474.64	302.38 (22.53)
<i>Anacardium occidentale</i>	25	44.928	47.693	46.446 (0.138)	0.103	0.32	0.161 (0.011)	9.2	57.9	146	441.34	375.79 (17.58)
<i>Eucalyptus grandis</i>	7	47.018	49.031	47.744 (0.350)	0.125	0.191	0.157 (0.011)	5.7	29.2	247.25	376.14	310.57 (21.94)

% dm, percentage of C and N in dry matter; n, number of selected trees. The stem wood samples of selected trees were extracted at 1.3 m of the ground. DBH range, range of diameter at breast height of sampled tree species. Figures in bracket represent the standard error (SE) of the mean

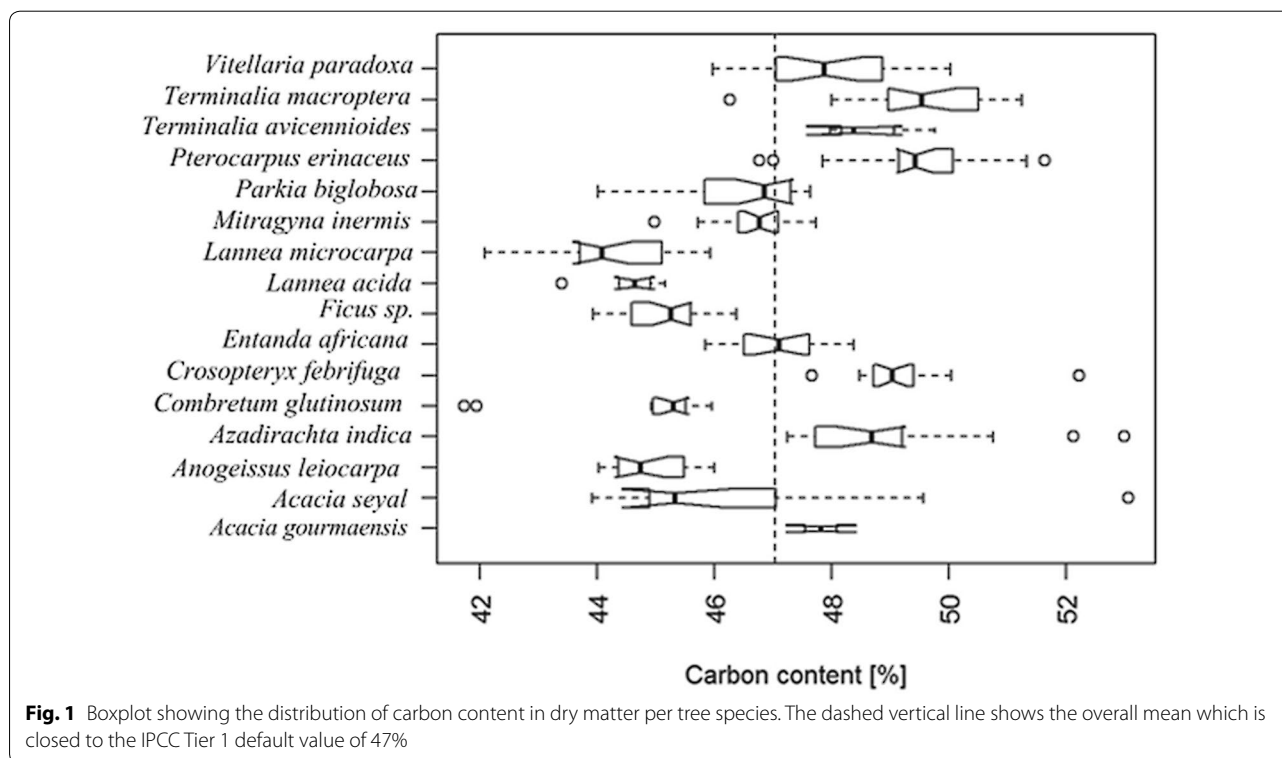


Fig. 1 Boxplot showing the distribution of carbon content in dry matter per tree species. The dashed vertical line shows the overall mean which is closed to the IPCC Tier 1 default value of 47%

the relevance of using in situ carbon content of the main tree species of the region (higher tier) for carbon accounting.

When applying the coefficient 0.5 as used by Chave et al. [8], Baccini et al. [4] to convert the mean biomass density into the mean carbon density for each LULC, the mean carbon density was overestimated for all LULC classes by 5.52% for Riparian forest and woodland, by 6.54% for Savannah Woodland, by 6.41% for Shrub Savannah, by 8.21% for grassland, by 7.6% for Cropland and Fallow, by 5.53% for Settlements, by 7.65% in Agroforestry systems and by 4.72% in Plantations. The application of the IPCC [42] default Tier 1 coefficient of 0.47 slightly overestimated carbon density by 0.15% (for Savannah Woodland), 0.54% (for Shrub Savannah), 1.72% (for grassland), 1.14% (for Cropland and Fallow), and 1.19% (for Agroforestry system) and underestimated by 0.81% (for Riparian forest and woodland), 0.80% (for Settlements) and 1.55% (for Plantation). We therefore recommend the use of the coefficient of 0.47 if one has to stick to the Tier 1 approach for carbon accounting in the Sudan Savannah environment.

The obtained carbon content for the most abundant species was in the same order of magnitude as results published by Guendehou et al. [37]; Andreae et al. [54]; Lasco et al. [55], Feldpausch et al. [56] and McGroddy et al. [57] even if the most abundant tree species varied considerably across the regions of the different case studies.

The nitrogen fraction of dry mater of the main tree species varied from 0.08% to 0.58%. The lowest mean nitrogen content for a single tree species was $0.128 \pm 0.012\%$ and the highest mean for a single tree species was $0.357 \pm 0.016\%$. The overall mean fraction of dry matter of nitrogen content was $0.229 \pm 0.016\%$. The species with highest nitrogen content in dry matter were *Acacia seyal*, *Acacia gourmensis*, *Ficus sp*, *Entanda Africana* and *Lannea microcarpa*. Human disturbance that affects these species could therefore lead to potentially high levels of N_2O emissions with high global warming potential due to the high fraction of nitrogen content into the dry matter of their stem wood. The C/N ratio per tree ranged from 80.71 to 570.05. The mean C/N per tree species ratio ranged from 135.97 ± 6.75 to 386.52 ± 28.28 for the different species for all land uses.

Carbon and nitrogen density and stocks at the landscape level

For the year 2013 the estimated stock in the watershed were for carbon $175,347.75 \pm 10,735.95$ Mg and for nitrogen 875.53 ± 51.76 Mg. The carbon density in $Mg C ha^{-1}$ were 44.81 ± 2.38 (for Riparian forest and woodland), 21.35 ± 1.16 (for Savannah Woodland), 6.57 ± 0.35 (for Shrub Savannah), 1.67 ± 0.15 (for Savannah grassland), 1.52 ± 0.14 (for Cropland and Fallow), 2.30 ± 0.48 (for Settlements), 21.39 ± 6.68 (for Agroforestry system) and 97.83 ± 27.55 (for Plantation) (Table 2). The carbon

Table 2 Mean carbon density ($Mg C ha^{-1}$) and total carbon stocks ($Mg C$) by LULC class at the watershed scale

LULC/LUCa	Descriptive statistic			Total carbon stock ($Mg C$) (SE)
	Range of carbon density ($Mg C ha^{-1}$)		Mean carbon density (SE)	
	Min	Max		
Forest land				$159,841.01 \pm 8721.48$
Riparian forest and woodland	35.46	57.27	44.81 (2.38)	$15,291.86 \pm 813.16$
Savannah Woodland	12.50	31.90	21.25 (1.16)	$116,401.70 \pm 6397.49$
Shrub Savannah	2.76	12.22	6.57 (0.35)	$28,147.43 \pm 1510.82$
Grassland				161.55 ± 15.23
Savannah grassland	0.03	2.98	1.67 (0.15)	161.55 ± 15.23
Cropland				$12,272.24 \pm 2326.92$
Cropland and Fallow	0.03	4.33	1.52 (0.14)	$12,272.24 \pm 2326.92$
Settlements				1125.66 ± 1187.20
Settlements	0.41	4.57	2.30 (0.48)	1125.66 ± 1187.20
Agroforestry				442.91 ± 138.47
Cashew plantation	4.99	98.08	21.39 (6.68)	442.91 ± 138.47
Plantation				1504.36 ± 435.29
<i>Eucalyptus grandis</i>	3.67	331.91	97.83 (27.55)	1346.27 ± 379.15
<i>Tectona grandis</i>	16.52	108.70	82.62 (33.09)	74.36 ± 29.78
<i>Azadirachta indica</i>	31.58	117.87	88.02 (28.23)	63.37 ± 0.32
<i>Gmelina arborea</i>	4.88	16.16	11.82 (3.50)	20.34 ± 6.02

The age of plantations and agroforestry system varied from 5 to 45 years old explained the large standard error (SE) and the large variance relative to the mean obtained from these plots. The area of each LULC class was provided in the Table 4

density was higher in settlements than in croplands and Savannah grasslands which is in line with our field observation that Biali community in this region tends to plant mostly trees species like *Azadirachta indica* within the settlements that are characterized by a high carbon density. Carbon density was higher in riparian forest and woodland than in cashew plantations. Both carbon content ($46.45 \pm 0.14\%$) and tree density (300 trees per ha) was much lower in cashew plantations (*Anacardium occidentale*) compared to riparian forests and woodlands (1397 trees per ha). This implies that the carbon offset when clearing a patch of riparian forest and woodland for farming activities unfortunately cannot be compensated by cashew plantations. We estimated this loss as $23.42 \text{ Mg C ha}^{-1}$. Despite the loss, it is important to adopt agroforestry after clearance of riparian forest since carbon loss is nearly twice as high for the conversion to cropland ($44.81 \pm 2.38 \text{ Mg C ha}^{-1}$). If Savannah woodland is converted to cashew plantations differences in mean carbon density are low while the conversion to cropland leads for both Savannah Woodland and for shrub Savannah to a net loss in carbon. Plantations with *Eucalyptus grandis*, *Tectona grandis*, *Azadirachta indica* had higher carbon densities per ha than riparian forests and could therefore be used to compensate carbon emissions from land clearing. The use of *Gmelina arborea* in

plantations compensates due to the low carbon density only partially for carbon emissions from land clearing.

For nitrogen (Table 3) relative effects of land use conversion were of similar magnitude as for carbon. Thus, the absolute differences are very different, but the relative differences are comparable. The different carbon and nitrogen density of the land use classes is reflected in the heterogeneous spatial distribution of carbon and nitrogen stocks at the watershed scale (Figs. 2, 3).

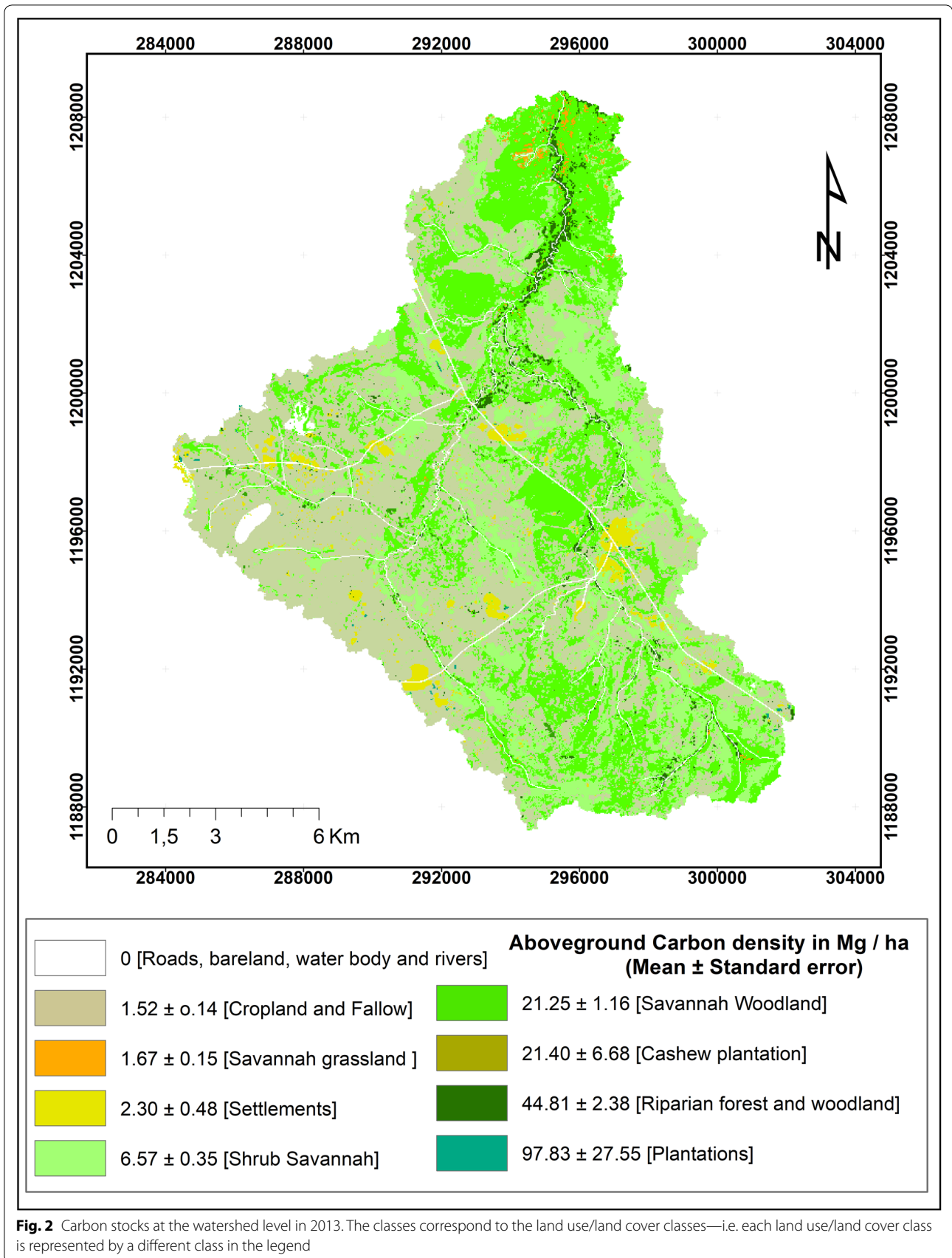
Conclusion

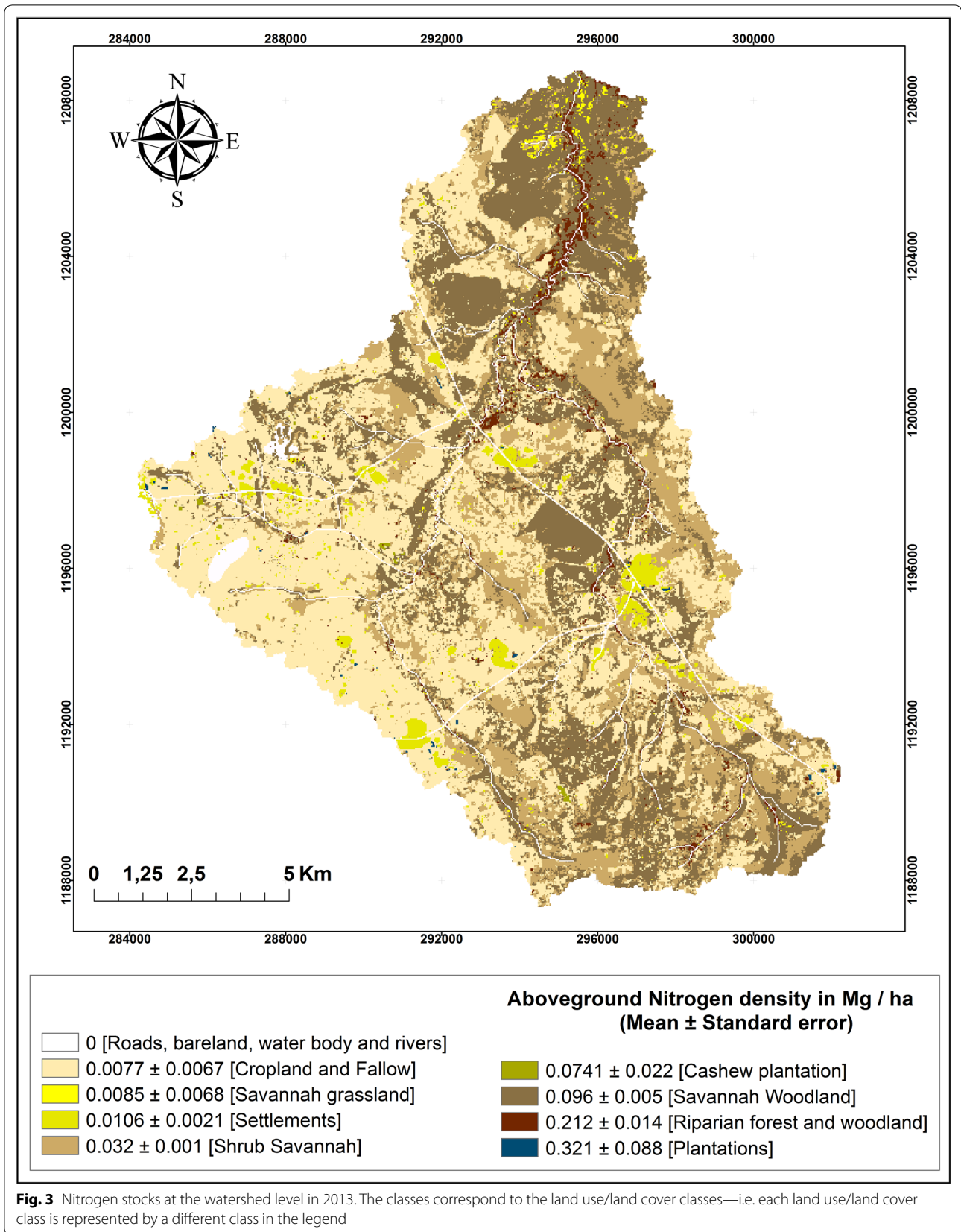
The results showed the relevance of using the in situ carbon and nitrogen content of the main tree species in estimating aboveground carbon and nitrogen stocks in the Sudan Savannah environment. By assessing the carbon and nitrogen fraction in dry matter of main tree species of the region uncertainty could be substantially reduced by 0.15 to 1.72% lower and 0.80 to 1.55% higher compared to the default IPCC [42] Tier 1 value of 47% depending the land use/land cover class. The overall mean carbon content across all land use categories as the average of 277 wood samples for all species was 47.01% indicating that a Tier 1 value 47% instead of the sometimes used value of 50% should be used in the Sudan Savannah environment if no more detailed information is present. Both results on carbon and nitrogen density in each LULC class, and

Table 3 Mean nitrogen density (Mg ha^{-1} of N) and total nitrogen stocks (Mg of N) by LULC class at the watershed scale

LULC/LUCa	Descriptive statistic			Total nitrogen stocks (Mg N) (SE)
	Range of nitrogen density (Mg ha^{-1} N)		Mean nitrogen density (Mg ha^{-1} N) (SE)	
	Min	Max		
Forest land				740.37 ± 0.021
Riparian forest and woodland	0.170	0.285	0.212 (0.014)	72.41 ± 0.014
Savannah Woodland	0.045	0.160	0.096 (0.005)	530.79 ± 0.005
Shrub Savannah	0.008	0.064	0.032 (0.001)	137.16 ± 0.001
Grassland				0.825 ± 0.0006
Savannah grassland	0.0001	0.0178	0.0085 (0.0068)	0.825 ± 0.00006
Cropland				62.57 ± 0.0006
Cropland and Fallow	0.00018	0.0252	0.0077 (0.0067)	62.57 ± 0.0006
Settlements				5.20 ± 0.002
Settlements	0.0017	0.0201	0.0106 (0.0021)	5.20 ± 0.002
Agroforestry				1.53 ± 0.022
Cashew plantation	0.017	0.340	0.0741 (0.022)	1.53 ± 0.022
Plantation				5.01 ± 0.0323
<i>Eucalyptus grandis</i>	0.012	1.091	0.321 (0.088)	4.42 ± 0.088
<i>Tectona grandis</i>	0.058	0.418	0.291 (0.115)	0.26 ± 0.115
<i>Azadirachta indica</i>	0.114	0.425	0.317 (0.101)	0.23 ± 0.101
<i>Gmelina arborea</i>	0.024	0.079	0.058 (0.017)	0.10 ± 0.017

The age of plantations and agroforestry system varied from 5 to 45 years which explained the large standard error (SE) and the large variance relative to the mean obtained from their plots data. The area of each LULC class was provided in the Table 1





the carbon and nitrogen content per trees species provide important information for carbon accounting related to the implementation of national REDD+ programmes of developing countries in the Sudan Savannah environment. Carbon stocks per ha in croplands and settlements in the case study region were comparable to Savannah grassland. Carbon stocks per ha in cashew plantations were comparable to Savannah Woodland but lower than riparian forests. The highest carbon stocks per ha were observed for plantations based on *Eucalyptus grandis*, *Tectona grandis*, or *Azadirachta indica*. While plantations of these three trees not endemic to West Africa are able to compensate carbon loss due to land use change trade-offs with other ecosystem goods and services and biodiversity should be considered.

Since the study took place at the local scale there is a need for the engagement of such work at the regional scale to confirm the importance of using in situ carbon and nitrogen data for carbon accounting. In this situation regional allometric equations are also of great importance for carbon accounting for the West African countries.

Materials and methods

Case study location

The region is located between 10°44'08" N–10° 55' 42"N and 1° 01' 32"E–1°11'30"E, specifically at the Dassari Basin situated in the North-West of Benin (Fig. 4) which a coverage area of 192.57 km². Long-term (1952–2010) minimal daily temperature ranged from 15.25 to 25.08 °C with an average of 20.53 °C. Daily maximum temperature ranged from 26.63 to 39.27 °C with a mean temperature of 32.59 °C. Long-term (1971–2013) mean annual rainfall was 1054.94 mm. The region was characterized by two periods of extreme droughts (1978–1979; 1985–1986) and some moderate to severe drought using the standardized precipitation index (SPI) programme developed by McKee [44].

Methods

Image classification

We coupled two scenes of Landsat 8 (<http://glovis.usgs.gov>) together with ground truthing information to classify land use/land cover. Landsat 8 satellite images from 13 October 2013 and 29 October 2013 were used—both with path-row 193-53. October was chosen since photosynthetic activity of natural vegetation and crops is high and cloud cover and fire pattern disturbance tend to be minimized during that part of the year.

Since it was not possible to separate agroforestry, forest land and plantations at the scale of the Landsat 8 data, these classes were separated based on several Worldview-2 (<http://www.digitalglobe.com>) imagery scenes with 0.5–2 m resolution together with additional ground

truth data from known agroforestry and plantation plots to discriminate agroforestry system and plantation from natural vegetation (cf. Fig. 5).

Based on the ground truthing data derived for the sample points (cf. Fig. 4), a random forest [46, 47] model was trained and used to classify the Landsat 8 data. The analysis was done in R [48] using the package random Forest [49]. The accuracy of the classification (Fig. 5) was acceptable to good as indicated by the overall accuracy of 0.75 and the kappa index of 0.70 [50].

Forest inventory

In reference to the objective of the current study we focused our measurements in the stand tree species of each LULC (land use land cover) of the site (Table 4). During the forest inventory we found some tree species such as *Vitellaria paradoxa*, *Parkia biglobosa*, *Lannea microcrapa* and *Lannea acida* which have the economic value for farmers and which were not burnt or cut off. The same remark is applicable to savannah grassland where we also have stand tree in low density. According to Zomer et al. [41] in sub-Saharan Africa, the majority (87%) of agriculturally dominated landscapes has a tree cover of more than 10%. For this purpose the measurements (DBH and Height) of stand tree species which are within crop land and fallow and savannah grassland are also of concerned like others LULC (Table 1) in this study.

Forest inventory was carried out from March to September 2014 in every LULC class. The plots were installed randomly proportionally to the area covered by the LULC class (Table 4) based on the equation of Pearson et al. [51]. The size of the plots was 30 m × 30 m in forest land, savannah grassland and cropland and fallow or agricultural land, 100 m × 100 m within settlements and 10 m × 20 m in agroforestry and plantation. A total number of 250 plots (Fig. 4 and Table 4) were surveyed—in total they covered 27.26 ha.

Importance Value Index (IVI) analysis

The IVI of a species is the sum of the relative frequency, relative density and relative dominance of the species [52]. Chabi et al. [45] estimated the IVI of the main species when developing biomass allometric models in the same watershed in the North-West of Benin. 84 species were identified during plots surveys. Three variables (DBH, total height of stand tree and wood density (Chabi et al. [45] of stem wood) were measured from each individual plant of DBH higher or equal than 5 cm. The identified main tree species were *Acacia seyal*, *Combretum glutinosum*, *Pterocarpus erinaceus*, *Anogeisus leiocarpus*, *Mitragyna inermis*, *Lannea microcrapa*, *Ficus sp*, *Crosoteryx febrifuga*, *Entada africana*, *Parkia biglobosa*, *Vitellaria paradoxa* and *Azadirachta indica* [45].

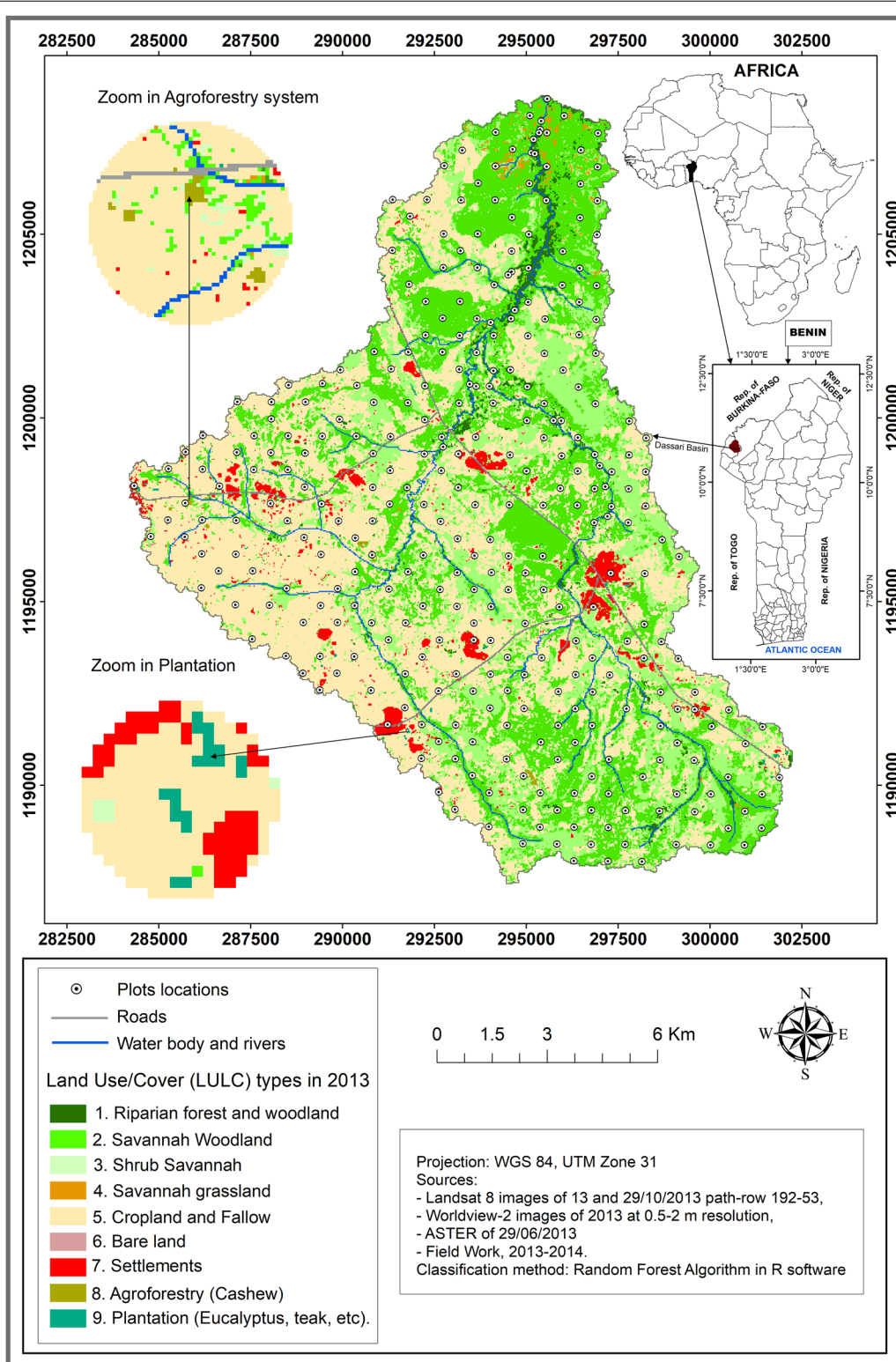


Fig. 4 Study area and land use/cover map of 2013/2014 with plots locations

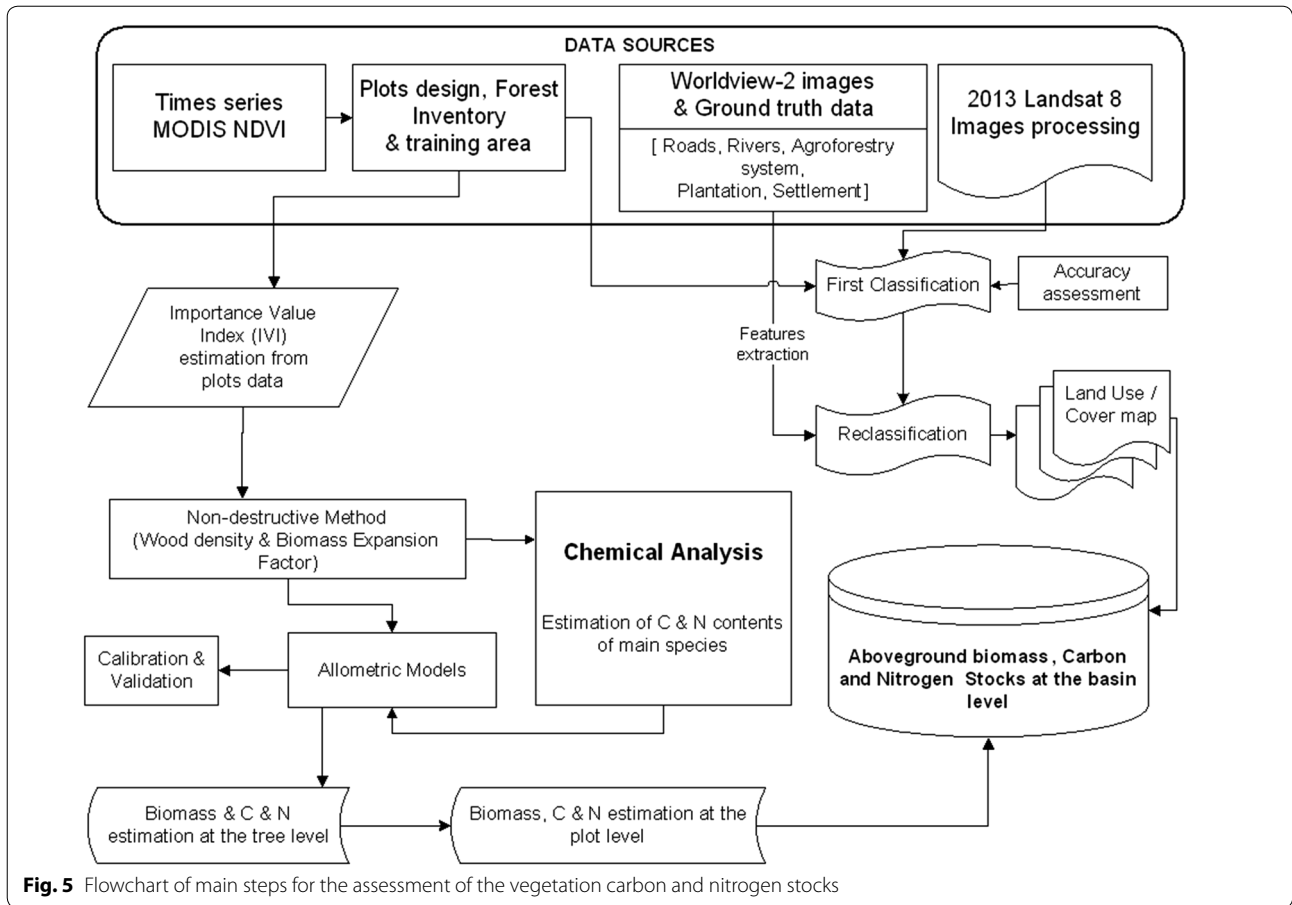


Table 4 Land use/land cover (LULC) classes and number of established plots

LUCa/LULC	Area (ha)	Percentage (%) in the basin	Area sampled (ha)	Number of established plots
Forestland				
Riparian forest and woodland	320.40	1.66	0.81	9
Savannah Woodland	5447.79	28.29	2.43	27
Shrub Savannah	4241.88	22.03	5.04	56
Grassland				
Savannah Grassland	96.48	0.50	3.06	34
Cropland				
Cropland and Fallow (Agricultural land)	8031.15	41.70	7.20	80
Settlement				
Settlement	486.72	2.53	8.00	8
Others land use				
Agroforestry	20.70	0.11	0.26	13
Plantation	16.74	0.09	0.46	23

NB: Agroforestry and plantation were identified as potential mitigation strategies to climate change. We therefore discriminated them from cropland. Percentage values do not add up to 100% since the basin also includes water bodies, roads and bare land that were not included in the table

Chemical analysis for the estimation of carbon and nitrogen content of stem wood samples

The main tree species of the different land use/land cover classes were identified based on tree inventory data derived during the first field trip. During the second field trip, stem wood samples from the main tree species were taken and analysed later on with respect to their carbon and nitrogen content. Additionally, diameter at breast height (DBH), tree height and wood density were assessed and used as input for an allometric model fitted to the local conditions [45].

During this second field trip, 277 stem wood samples from 18 tree species were obtained. After wood density estimation samples were re-dried, grinded and weighted. Chemical analysis was done at the Institute of Crop Science and Resource Conservation, within the laboratory of the Department of Plant Nutrition in Germany (Bonn) using the EA3000 model CHNS-O Elemental Analyser (<http://www.eurovector.it/>).

Assessment of aboveground carbon and nitrogen stocks

The methodological approach to calculate the carbon and nitrogen stocks was similar across all LULC of Table 4.

For this purpose as only stand trees species were concerned in this study, the estimation of aboveground carbon and nitrogen stocks was based on the biomass estimation at the tree level using the published equations from Chabi et al. [45] corresponding to each LULC for all tree species, except for two tree species. For the Senegal date palm (*Phoenix reclinata*) and the Asian Palmyra palm (*Borassus flabellifer*) biomass was estimated using the equation from Schroth [53] developed for coconut tree (*Cocos nucifera*) which is a member of the family Areaceae (palms) such as *Borassus flabellifer* and *Phoenix reclinata*. For the estimation of aboveground biomass of tree species of crop land and fallow and the savannah grassland we also apply the published equations from Chabi et al. [45] corresponding to these two LULC classes. These published equations can be found in the additional file 2 of Chabi et al. [45].

By combining the carbon content of the different tree species or the nitrogen content of the different tree species (Table 1) with the biomass estimated from the allometric models Chabi et al. [45], carbon and nitrogen stocks were estimated at the tree and the plot level (Eqs. 1a; 2a, 3 and 4). When the tree species did not belong to the main tree species of Table 1, we applied the overall mean of carbon and nitrogen content across all species to estimate their carbon and nitrogen stocks (Eqs. 1b, 2b).

$$C_t = C_{ts} * B_t \quad (1a)$$

$$C_t = C_{mc} * B_t \quad (1b)$$

$$N_t = N_{ts} * B_t \quad (2a)$$

$$N_t = N_{mn} * B_t \quad (2b)$$

$$C_p = \sum_{i=1}^n C_{ti} \quad (3)$$

$$N_p = \sum_{i=1}^n N_{ti} \quad (4)$$

where: B_t , Biomass at the tree level and this is the function of the published equation from Chabi et al. [45]; C_t , The carbon stock in the dry matter at the tree level; C_{ts} , The fraction of Carbon content of the tree species or the percentage of C in the dry matter of the tree species; C_{mc} , The mean fraction of carbon content for all 277 wood samples in the case study. C_{mc} equal to 0.4701. The IPCC [42] default value is equal to 0.47. C_{mc} is used when the tree species did not belong to the tree species of the Table 1; N_t , The nitrogen stock in the dry matter at the tree level; N_{ts} , The fraction of Nitrogen content of the tree species or the percentage of N in the dry matter of the tree species; N_{mn} , The mean fraction of nitrogen content for all 277 wood samples in the case study. N_{mn} equal to 0.229; C_p , The carbon stock at the plot level; N_p , The nitrogen stock at the plot level; n , The total number of tree species in the plot, the index variable i goes from 1 to n .

By combining information from carbon and nitrogen stocks at plot level with the land use/land cover classification (Table 4), carbon as well as nitrogen stocks for each LULC were calculated as mean carbon and nitrogen density (Eqs. 5 and 6), (Tables 2 and 3) times the area of the LULC class (Table 4 and Fig. 5).

$$C_{dLULC} = \frac{\sum_{i=1}^{np} C_{pi}}{np} \pm \varepsilon \quad (5)$$

$$N_{dLULC} = \frac{\sum_{i=1}^{np} N_{pi}}{np} \pm \varepsilon \quad (6)$$

where: C_{dLULC} , Carbon density for each LULC expressed in Mg/ha with associated standard error (ε); N_{dLULC} , Nitrogen density for each LULC expressed in Mg of N per ha with associated standard error (ε); np , The total number of the plots in each LULC, the index variable i goes from 1 to np ; C_{pi} , The carbon stock of the plot i ; N_{pi} , The nitrogen stock of the plot i .

The carbon and nitrogen stocks maps were compiled in ArcGIS 10.2.1 (<http://www.esri.com/>) and visualized (Figs. 2 and 3).

Abbreviations

C: carbon; DBH: diameter at breast height; IVI : Importance Value Index; IPCC: Intergovernmental Panel on Climate Change; LUCa: land use category; LULC: land use/land cover; Mg: megagramme; N: nitrogen; REDD+: reducing emissions from deforestation and forest degradation, biodiversity conservation, sustainable forest management and enhancement of forest carbon stocks; SE: standard error; SPI: standardized precipitation index.

Acknowledgements

The authors are most grateful to the German Federal Ministry of Education and Research (BMBF) for sponsorship. As the corresponding author, I am sincerely grateful to Prof. Dr. Sven Lautenbach, who provided 60% of the cost of the chemical analysis for carbon and nitrogen estimation of stem wood samples. We are also grateful to INRES (Bonn-Germany) staff that did the chemical analysis.

I also thank NASA (National Aeronautics and Space Administration) for granting access to Landsat 8 images.

Authors' contributions

AC provided literature indications, designed the methodology, carried out data collection and drafted the manuscript. SL designed the methodology, performed statistical analysis and revised the manuscript. JET, VAO and SAB provided methodology and revised the manuscript. NKB supervised the research and revised the manuscript. JVM and JF revised the manuscript. All authors read and approved the final manuscript.

Authors' information

AC is a part-time lecturer at the University of Parakou in the Republic of Benin and part-time junior scientist at WASCAL competence centre. SL is a senior scientist at the GIScience Research Group at the Heidelberg University. JET is the Director of Research at WASCAL competence centre. VAO is Associate Professor at the University of Abomey-Calavi in Benin. SAB is the senior scientist at the forestry research institute of Ghana. NKB is Professor at Kwame Nkrumah University of Science and Technology in Ghana and member of CCLU (Climate Change and Land Use). JVM is a senior scientist at the National Institute for Agricultural Research of Benin (INRAB). JF is forestry officer in REDD + programme at FAO Ghana Regional office.

Funding

Not applicable.

Availability of data and materials

We declare available data and material used in the setting of this study.

Ethics approval and consent to participate

All authors were involved in this study from the beginning to the end.

Consent for publication

All authors have consented for publication.

Competing interests

The authors declare that they have no competing interests.

Author details

¹ Department of Geography, University of Parakou, Parakou, Republic of Benin. ² West African Science Service Centre On Climate Change and Adapted Land Use (WASCAL), Competence Centre Ouagadougou, 06 BP 9507 Ouaga 06 Ouagadougou, Burkina Faso. ³ GIScience Research Group, Institute of Geography, University of Heidelberg, Mathematikon, Berliner Str. 45, 4. OG, Raum 004, Heidelberg, Germany. ⁴ Institute of Geodesy and Geoinformation, University of Bonn, Nussallee 1, 53115 Bonn, Germany. ⁵ Department of Geography, University of Abomey-Calavi, BP 677 Abomey-Calavi, Republic of Benin. ⁶ CSIR-Forestry Research Institute of Ghana, Tryft, University P.O. Box 63, Kumasi, Ghana. ⁷ Departments of Agricultural Engineering, The College of Engineering, Kwame Nkrumah University of Science and Technology, Kumasi, Ghana. ⁸ National Institute for Agricultural Research of Benin (INRAB), 06 BP 1105 Cotonou, Benin. ⁹ Food & Agriculture Organization of the United Nations (FAO), Regional Office for Africa (RAF), Accra, Ghana.

Received: 8 October 2018 Accepted: 30 August 2019

Published online: 10 September 2019

References

- Angelsen, Wertz-Kanounnikoff S. Moving ahead with REDD: issues, options and implications. Bogor. 2008. https://www.cifor.org/publications/pdf_files/Books/BAngelsen0801.pdf. Accessed 31 Aug 2018.
- GOF-C-GOLD. A sourcebook of methods and procedures for monitoring and reporting anthropogenic greenhouse gas emissions and removals associated with deforestation, gains and losses of carbon stocks in forests remaining forests, and forestation. GOF-C-GOLD Rep. version COP21. 2015. http://www.gofc-gold.uni-jena.de/redd/sourcebook/Sourcebook_Versi_on_Nov_2010_cop16-1.pdf.
- Houghton RA, Hole W, Hole W. Aboveground forest biomass and the global carbon balance. *Glob Chang Biol*. 2005;11:945–58.
- Baccini A, Goetz SJ, Walker WS, Laporte NT, Sun M, Hackler J, et al. Estimated carbon dioxide emissions from tropical deforestation improved by carbon-density maps. *Nat Clim Chang*. 2012;2:182–5. <https://doi.org/10.1038/nclimate1354>.
- Defries RS, Houghton RA, Hansen MC, Field CB, Skole D, Townshend J. Carbon emissions from tropical deforestation and regrowth based on satellite observations for the 1980s and 1990s. *Proc Nat Acad Sci* 2002;99(22):14256–61. <https://doi.org/10.1073/pnas.182560099>.
- Grassi G, Monni S, Federici S, Achard F, Mollicone D. Applying the conservativeness principle to REDD to deal with the uncertainties of the estimates. *Environ Res Lett*. 2008;3:035005. <https://doi.org/10.1088/1748-9326/3/3/035005>.
- Pelletier J, Ramankutty N, Potvin C. Diagnosing the uncertainty and detectability of emission reductions for REDD + under current capabilities: an example for Panama. *Environ Res Lett*. 2011;6:024005. <https://doi.org/10.1088/1748-9326/6/2/024005>.
- Chave J, Réjou-Méchain M, Búrquez A, Chidumayo E, Colgan MS, Delitti WBC, et al. Improved allometric models to estimate the aboveground biomass of tropical trees. *Glob Chang Biol*. 2014;20:3177–90.
- Ter-mikaelian MT, Korzukhin MD. Biomass equations for sixty-five North American tree species. *For Ecol Manage*. 1997;97:1–24.
- Jose JJS. Carbon stocks and fluxes in a temporal scaling from a savanna to a semi-deciduous forest. *For Ecol Manage*. 1998;105:251–62.
- Nelson BW, Mesquita R, Pereira JLG, De SGA, Teixeira G, Bovino L. Allometric regressions for improved estimate of secondary forest biomass in the central Amazon. *For Ecol Manage*. 1999;117:149–67.
- Clark DB, Clark DA. Landscape-scale variation in forest structure and biomass in a tropical rain forest. *For Ecol Manage*. 2000;137:185–98.
- Grierson P, Williams K. Review of unpublished biomass-related information: Western Australia. New South Wales and Queensland: South Australia; 2000.
- Keith H, Barrett D, Keenan R. Review of Allometric Relationships for Estimating Woody Biomass for New South Wales, the Australian Capital Territory, Victoria, Tasmania and The National Carbon Accounting System: national carbon accounting system technical report no. 5b. *Soil Sci Soc Am J*. 2009;73:2078. <https://www.soils.org/publications/sssaj/abstracts/73/6/2078>.
- Keller M, Palace M, Hurtt G. Biomass estimation in the Tapajos National Forest, Brazil Examination of sampling and allometric uncertainties. *For Ecol Manage*. 2001;154(154):371–82.
- Fleurant C, Duchesne J, Raimbault P. An allometric model for trees. *J Theor Biol*. 2004;227:137–47.
- Jenkins JC, Chojnacky DC, Heath LS, Birdsey RA. Comprehensive database of diameter-based biomass regressions for North American Tree species. 2004.
- Zianis D, Mencuccini M. On simplifying allometric analyses of forest biomass. *For Ecol Manage*. 2004;187:311–32.
- Domke GM, Woodall CW, Smith JE, Westfall JA, Mcroberts RE. Forest ecology and management consequences of alternative tree-level biomass estimation procedures on US forest carbon stock estimates. *For Ecol Manage*. 2012;270:108–16. <https://doi.org/10.1016/j.foreco.2012.01.022>.

20. Akindele SO, Lemay VM. Development of tree volume equations for common timber species in the tropical rain forest area of Nigeria. *For Ecol Manage.* 2006;226:41–8.
21. Dossa EL, Fernandes ECM, Reid WS, Ezui K. Above- and belowground biomass, nutrient and carbon stocks contrasting an open-grown and a shaded coffee plantation. *Agro For Syst.* 2008;72:103–15.
22. Djomo AN, Ibrahima A, Saborowski J, Gravenhorst G. Allometric equations for biomass estimations in Cameroon and pan moist tropical equations including biomass data from Africa. *For Ecol Manage.* 2010;260:1873–85. <https://doi.org/10.1016/j.foreco.2010.08.034>.
23. Djuikouo MNK, Jean-Louis D, Nguembou CK, Lewis SL, Sonke B. Diversity and aboveground biomass in three tropical forest types in the Dja Biosphere Reserve, Cameroon. *Afr J Ecol.* 2010;48:1053–63.
24. Sawadogo L, Savadogo P, Tiveau D, Djibril S, Zida D, Nouvellet Y, et al. Allometric prediction of above-ground biomass of eleven woody tree species in the Sudanian savanna-woodland of West Africa. *J For Res.* 2010;21:475–81.
25. Henry M, Picard N, Trotta C, Manlay RJ, Valentini R, Bernoux M, et al. Estimating tree biomass of Sub-Saharan African forests: a review of available allometric equations. *SILVA Fenn.* 2011;45(3B):477–569.
26. Rasmussen MO, Göttsche F-M, Diop D, Mbow C, Olesen F-S, Fensholt R, et al. Tree survey and allometric models for tiger bush in northern Senegal and comparison with tree parameters derived from high resolution satellite data. *Int J Appl Earth Obs Geoinf.* 2011;13:517–27. <https://doi.org/10.1016/j.jag.2011.01.007>.
27. Shirima DD, Munishi PKT, Lewis SL, Burgess ND, Marshall AR, Balmford A, et al. Carbon storage, structure and composition of miombo woodlands in Tanzania's Eastern Arc Mountains. *Afr J Ecol.* 2011;49:332–42.
28. Bakayoko O, Assa AM, Coulibaly B, N'Guessan KA. Stockage de Carbone dans des peuplements de *Cedrela Odorata* et de *Gmelina Arborea* en Côte d'Ivoire. *Eur J Sci Res.* 2012;75:490–501. <http://www.europeanjournalofscientificresearch.com>.
29. Kuyah S, Dietz J, Muthuri C, Jamnadass R, Mwangi P, Coe R, et al. Agriculture, ecosystems and environment allometric equations for estimating biomass in agricultural landscapes: i. Aboveground biomass. *Agricult Ecosyst Environ.* 2012;158:216–24. <https://doi.org/10.1016/j.agee.2012.05.011>.
30. Mbow C, Verstraete MM, Sambou B, Tahirou A, Neufeldt H. Allometric models for aboveground biomass in dry savanna trees of the Sudan and Sudan–Guinean ecosystems of Southern Senegal. *J For Res.* 2014;19:340. <https://doi.org/10.1007/s10310-013-0414-1>.
31. Ngomanda A, Laurier N, Obiang E, Lebamba J, Moundounga Q, Gomat H, et al. Site-specific versus pantropical allometric equations: Which option to estimate the biomass of a moist central African forest? *For Ecol Manage.* 2014;312:1–9. <https://doi.org/10.1016/j.foreco.2013.10.029>.
32. Li C, Xiao C. Above- and belowground biomass of *Artemisia ordosica* communities in three contrasting habitats of the Mu Us desert, northern China. *J Arid Environ.* 2007;70:195–207.
33. Basuki TM, Van Laake PE, Skidmore AK, Hussin YA. Allometric equations for estimating the above-ground biomass in tropical lowland Dipterocarp forests. *For Ecol Manag J.* 2009;257:1684–94.
34. Fonton NH, Yabi CC, Dah-dovonon JZ. Modélisation du volume du fût d'arbre pour une gestion durable des écosystèmes forestiers soudanais. 2009.
35. Navar-Chaidez JDJ. Biomass allometry for tree species of northwestern Mexico. *Trop Subtrop Agroecosystems.* 2010;12:507–19. <http://redalyc.uaemex.mx/src/inicio/ArtPdfRed.jsp?iCve=93915170011>.
36. Morote FAG, Serrano FRL, Andrés M, Rubio E, Jiménez JLG, De Heras J. Allometries, biomass stocks and biomass allocation in the thermophilic Spanish juniper woodlands of Southern Spain. *For Ecol Manage.* 2012;270:85–93.
37. Guendehou GHS, Lehtonen A, Moudachirou M, Mäkipää R, Sinsin B. Stem biomass and volume models of selected tropical tree species in West Africa. *South For.* 2012;74:77–88.
38. Aholoukpè H, Dubos B, Flori A, Deleporte P, Amadjì G, Chotte JL, et al. Estimating aboveground biomass of oil palm: allometric equations for estimating frond biomass. *For Ecol Manage.* 2013;292:122–9.
39. Hunter MO, Keller M, Victoria D, Morton DC. Tree height and tropical forest biomass estimation. *Biogeosciences.* 2013;10:8385–99.
40. Montagnoli A, Fusco S, Pflugmacher D, Cohen WB, Scippa GS. Forest Ecosystems Estimating forest aboveground biomass by low density lidar data in mixed broad-leaved forests in the Italian Pre-Alps. 2015.
41. Zomer RA, Trabucco A, Coe R, Place F. Trees on farm: analysis of global extent and geographical patterns of Agroforestry ICRAF Working Paper no. 89. 2009; <http://www.worldagroforestry.org/downloads/publications/PDFs/WP16263.PDF>.
42. IPCC. 2006 IPCC Guidelines for National Greenhouse Inventories—a primer, Prepared by the National Greenhouse Gas Inventories Programme, Intergov Panel Clim Chang Natl Greenh Gas Invent Program. 2008; 20.
43. Chave J, Andalo C, Brown S, Cairns MA, Chambers J, Eamus D, et al. Tree allometry and improved estimation of carbon stocks and balance in tropical forests. *Oecologia.* 2005;145:87–99.
44. Mckee TB, Doesken NJ, Kleist J. The relationship of drought frequency and duration to time scales. In: 8th Conference on Applied Climatology, Anaheim, 17–22 January 1993, p. 17–22.
45. Chabi A, Lautenbach S, Orekan VOA, Kyei-Baffour N. Erratum : Allometric models and aboveground biomass stocks of a West African Sudan Savannah watershed in Benin. [*Carbon Balance Manage*, 11, (2016), (16)] <https://doi.org/10.1186/s13021-016-0058-5>. *Carbon Balance Manag.* 2016;11.
46. Breiman L. Random forests. *Mach Learn.* 2001;45:5–32. <https://doi.org/10.1023/A:1010933404324>.
47. Gislason PO, Benediktsson JA, Sveinsson JR. Random forests for land cover classification. *Pattern Recognit Lett.* 2006;27:294–300.
48. R Core Team. A language and environment for statistical computing. Computing. 2006.
49. Liaw A, Wiener M. Classification and regression by random forest. *R news.* 2002;2:18–22.
50. Foody GM. Status of land cover classification accuracy assessment. *Remote Sens Environ.* 2002;80:185–201.
51. Pearson TRH, Brown SL, Birdsey RA. Measurement guidelines for the sequestration of forest carbon. 2007.
52. Curtis JT, Cottam G. The use of distance measures in phytosociological sampling author(s): Grant Cottam and J. T. Curtis Reviewed work (s): Published by : Ecological Society of America Stable. <http://www.jstor.org/stable/1930167>. America (NY). 2012;37:451–60.
53. Schroth G, D'Angelo SA, Teixeira WG, Haag D, Lieberei R. Conversion of secondary forest into agroforestry and monoculture plantations in Amazonia: consequences for biomass, litter and soil carbon stocks after 7 years. *For Ecol Manage.* 2002;163:131–50.
54. Andreae MO, Merlet P. Emission of trace gases and aerosols from biomass burning. *Global Biogeochem Cycles.* 2001;15:955–66.
55. Lasco RD, Pulhin FB. Philippine forest ecosystems and climate change: carbon stocks, rate of sequestration and the Kyoto protocol. *Ann Trop Res.* 2003;25:37–51.
56. Internacional C, Tropical DA, Sciences A, Amazo E. Carbon and nutrient accumulation in secondary forest regenerating on pastures in Central Amazonia. *Ecol Appl.* 2004;14:164–76.
57. Mcgroddy ME, Daufresne T, Hedin LO. Scaling of C: n: P Stoichiometry in forest worldwide: Implication of terrestrial redfield-type ratios special feature. *Ecol Soc Am Scaling.* 2004;85:2390–401.
58. Chaturvedi RK, Raghubanshi AS. Aboveground biomass estimation of small diameter woody species of tropical dry forest. *New Forest.* 2013;2013(4):509–19.
59. Chaturvedi RK, Raghubanshi AS. Assessment of carbon density and accumulation in mono- and multi-specific stands in Teak and Sal forests of a tropical dry region in India. *For Ecol Manage.* 2015;339:11–21. <https://doi.org/10.1016/j.foreco.2014.12.002>.
60. Chaturvedi RK, Raghubanshi AS. Allometric models for accurate estimation of aboveground biomass of teak in tropical dry forests of India. *For Sci.* 2015;61:938–49.

Publisher's Note

Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.